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From bottom to up: effects of fishery improvement projects on the stock status of multi-specific small-scale fisheries from Mexico --Manuscript Draft--

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Dear Editor,
Marine policy

We wish to submit an original research article entitled “**From bottom to up: effects of fishery improvement projects on the stock status of multi-specific small-scale fisheries from Mexico.**” for consideration for publication in *Marine Policy*. We confirm that this work is original and has not been published elsewhere, nor is it currently under consideration for publication elsewhere.

In this paper, we evaluated the current status of ten species involved in Fishery Improvement Projects (FIPs) in Mexico using data-poor methods. To make these assessments we had to reconstruct historical catches for these species. We also discussed the implications of the catches of communities involved in these FIPs for the total catch of the stock and the potential utility of these projects to aid in filling data gaps needed to assess the status of exploited stocks.

Our results are relevant because there are no species-specific catch data nor stock assessments in Mexico for the majority of species and, like in other data-poor areas, this situation prevents to understand the status of fishing resources and how FIPs contribute to their sustainable use. In addition, we believe that the results obtained in this study could serve as the basis to define management actions that helps to secure their sustainable use and to identify data gaps needed to implement more robust assessments.

We believe that this manuscript is appropriate for publication by Marine Policy because it contributes to a better understanding of the status of fishing resources in data-poor areas, highlights the importance of fishing communities to fill data gaps, and the results could inform stakeholders on how to define proper management actions for exploited stocks.

We also declare that we have no conflicts of interest to disclose. All sources of funding are acknowledged in the manuscript.

Please address all correspondence concerning this manuscript to me at ososa@cicese.mx. Thank you for your consideration of this manuscript. We appreciate your time and look forward to your response

Sincerely,
Oscar Sosa-Nishizaki

Abstract

Small-scale fisheries are socioeconomically important in data poor areas as a source of income, food, and employment. However, most fisheries lack information (e.g., catch and effort) needed to evaluate their status, and are poorly managed. Fishery Improvement Projects (FIP) have emerged as a community-driven option to enhance management since they require to evaluate the status of harvested stocks. We assessed the stock status of ten species involved in FIPs in Mexico, using data-poor methods based on catches. Due to the historical inconsistencies of landing reports in the country, data reconstructions have to be made for most of the stocks evaluated. Data generated by communities involved in FIPs were used to refine these reconstructions and to inform assessment models. Results showed that most of the stocks are fished at unsustainable levels which may be related to increases in landings for the last ten years. Models including abundances indexes, estimated from FIPs data, produced better estimations. Stock status indicate that management actions area needed along with improvements in data collection. Improvement projects represent an opportunity to fill information gaps and inform assessment models in data-poor areas when official data is not available, which will help to secure the sustainability of small-scale fisheries.

From bottom to up: effects of fishery improvement projects on the stock status of multi-specific small-scale fisheries from Mexico

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5 specific small-scale fisheries from Mexico
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7 8 **Abstract** 9

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12 income, food, and employment. However, most fisheries lack information (e.g., catch and
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25 official data is not available, which will help to secure the sustainability of small-scale
26 fisheries.
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34 **1. Introduction** 35

36 Small-scale fisheries contribute to about half of the global fish catches and represent two-
37 thirds of the yields directed to human consumption [1]. However, the information on their
38 operations and catch is only sometimes available. This data scarcity leads to small-scale
39 fisheries being often overlooked and unassessed. Around 80% of global fish catches come
40 from unassessed fisheries, including most small-scale fisheries [2]. Furthermore, recent
41 analysis suggests that marine fish stocks, with an assessment of their status, are in better
42 shape and with effective management systems [3], highlighting the need to evaluate the
43 stocks exploited by small-scale fisheries to improve their management. In 2019, 64.6% of
44 assessed stocks by FAO were within biologically sustainable levels, while 35.4% were at
45 biologically unsustainable levels [4].
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49 In Mexico, coastal finfish fisheries (locally known as “escama”) have a significant
50 socioeconomic impact on many coastal communities as a source of income, food, and
51 employment [5–7]. Finfishes are caught by small-scale vessels (<10 m) using hooks, gillnets
52 (bottom and surface set), longlines (bottom and surface set), and traps [6,8,9]. These fisheries
53 are multi-specific; their catches include more than 440 bony fishes categorized in 17 groups
54 [5,9,10].
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57 Finfish fisheries in Mexico are managed by the National Commission of Fisheries and
58 Aquaculture (CONAPESCA) and the National Institute of Fisheries and Aquaculture
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4 (INAPESCA) which oversees the collection and publication of landings in the Fisheries
5 Statistics Yearbooks, the publication of the National Fisheries Charter (NFC), which is one
6 of the leading management tools where fisheries management reference points and
7 recommendations are published [11], and the developing the Fisheries Management Plans
8 for each fishery. Despite the fishing importance of these species, the deficiencies in the
9 official catch records (lack of catches time series by species and null reports of fishing effort)
10 and the scarce data on their biology make their populations assessments and sustainable
11 management extremely difficult [12,13].
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15 Since the 2000s, several small-scale fisheries worldwide began to record more specific own
16 information with stakeholders' support, favoring their fisheries' co-management and
17 promoting cooperative research [14]. At the same time, the development of environmentally
18 responsible fishing standards has been highlighted through the certification of fisheries and
19 recommendations for the consumption of seafood [15,16] and gave rise to the Fishery
20 Improvement Projects (FIPs) in 2002 [17].
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24 FIPs aim to develop sustainable fishing practices and represent a great opportunity and frame
25 of reference for improving resource management in developing countries like Mexico. In this
26 country, 97% of the total fishing effort (total = 76,880 fishing vessels) is operated by small-
27 scale fisheries that are poorly managed [18]. Further, FIPs represent an opportunity to
28 produce robust species-specific information to aid in assessing and managing exploited
29 stocks. Currently, 37 FIPs are implemented in Mexico (15% of all FIPs worldwide) [19].
30 Nine FIPs in Mexico focus on finfish fishery species, representing a significant collaborative
31 effort between stakeholders and citizen science toward the sustainability of these data-limited
32 small-scale fisheries [13].
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36 During stage one of the FIP development, the assessment of the fish stock status of the fishery
37 is needed to be associated with an accurate indicator of its level [20]. Stock assessments aim
38 to understand how many fish can be sustainably caught by fitting population dynamics
39 models to fisheries monitoring data. While comparing with biological reference points, they
40 allow for assessing whether the stock is in an overfishing state or overfished [21] and guide
41 operational fisheries management actions. Depending on the available data, several
42 quantitative approaches have been developed to assess fisheries stocks using data-limited to
43 rich and intensive data and costly methodologies [22]. For most small-scale fisheries, the
44 need for more data-rich information to perform robust stock assessments is common [23,24].
45 Therefore, data-limited stock assessment methods have been developed to fill these gaps
46 [2,25,26].
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50 Within the data-limited stock assessments methods, catch-only methods provide a simple
51 approach to produce estimates for stock biomass status (B/B_{msy}) and other reference points
52 based primarily on the trends of catch or landings time series and life history traits [27–29].
53 Among these methods, Catch-MSY (CMSY) is a mechanistic method developed by Martell
54 and Froese [25] and improved by Froese et al. [30] that has been used widely to determine
55 the stock status of global [31,32] and regional fisheries [33,34]. This method has been used
56 in Mexico to evaluate the stock status of the Pacific angel shark (*Squatina californica*; [35])
57 and the barred sand bass (*Paralabrax nebulifer*), which later led to the implementation of a
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4 fishery management plan [36]. Due to the deficiencies in the fishing report system in Mexico
5 described above, these methods pose an opportunity to give insights into the stock status of
6 fished species and to develop proper management actions that secure the sustainable use of
7 fishing resources.
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10 In this study, we used the information produced, up to date, from selected cases of small-
11 scale finfish fisheries associated with five FIPs to estimate the stock status of nine fish stocks
12 from Northwestern Mexico and one in the Gulf of Mexico using the CMSY method. Then,
13 we discuss the challenges that the development of FIPs confronts under the Mexican fisheries
14 management systems and their potential to improve this system from the bottom up.
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20 **2. Methods**

21 **2.1. Study cases**

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23 We assessed the status of ten finfish stocks in coastal areas from the Mexican Pacific
24 Northwest and the Gulf of Mexico (Fig. 1). The studied cases included the Ocean whitefish
25 (*Caulolatilus princeps*), California sheephead (*Semicossyphus pulcher*), barred sand bass
26 (*Paralabrax nebulifer*), vermilion rockfish (*Sebastes miniatus*), starry rockfish (*S.*
27 *constellatus*), yellowtail (*Seriola lalandi*), rooster hind (*Hyporthodus acanthistius*),
28 goldspotted sand bass (*Paralabrax auroguttatus*), Pacific red snapper (*Lutjanus peru*), and
29 the northern red snapper (*Lutjanus campechanus*) (Table 1). These coastal fishes have a
30 variety of life-history traits, with maximum ages ranging between 20 to 61 years [37–39] and
31 age at maturity ranging from 2 to 6 years [38,40,41]. Most of the species are broadcast
32 spawners, except for the *Sebastes* species, in which the fertilization and embryo development
33 occur internally [40]. In addition, *S. pulcher* is a protogynous hermaphrodite species that can
34 transition from a reproductively functional female to a male [37,38].
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Figure 1. Location of the communities where the Fishery Improvement Projects are being implemented in northwestern Mexico and the Gulf of Mexico.

Table 1. Study species from five different Fishery Improvement Projects (FIP) in Mexico. The geographic locations of the FIPs are shown in Fig. 1.

FIP	Species	Fishing gear	Start of the FIP	Number of vessels
Guaymas	<i>S. lalandi</i> , <i>L. peru</i> , <i>C. princeps</i> , <i>P. auroguttatus</i> , <i>H. acanthistius</i>	Handline	2018	16
El Rosario	<i>C. princeps</i> , <i>P. nebulifer</i> , <i>S. pulcher</i> , <i>S. miniatus</i> , <i>S. constellatus</i>	Handline and trap	2018	20
Natividad Island	<i>C. princeps</i>	Handline and trap	2017	18

San Cosme-Punta Coyote Corridor	<i>L. peru, S. lalandi</i>	Handline	2019	29
Nuevo Campechito	<i>L. campechanus</i>	Bottom and vertical longlines	2018	6

2.1. Data sources

Details of data sources used for landing reconstructions and assessment of stock status are reported in Table 2. The baseline for this analysis was the official landing statistics reported in the CONAPESCA’s Fisheries Statistics Yearbook. Formerly, other agencies have produced these statistics, causing differences in the amount and quality of the information reported. Yearbooks contain landings by common name (that could include multiple species), by Fishing Office and State when most detailed. Landings reported from 1970 to 1999 were used. Since 2000, more detailed landing statistics have been made available through the National Transparency Platform. From 2000 to 2019 (last year with official landings records), landings from this source were used. All landings used for these analyses were “landed weight,” as fishers reported this information directly.

Table 2. Data sources available for the assessment of ten stocks in Mexico.

Species	Catch period	Abundance indexes	Catch reconstructed	Base for the reconstruction
<i>S. lalandi</i>	1970-2019	CPUE (2011-2019)*	Yes	Limit of distribution (Baxter 1960, Ulloa-Ramírez et al. 2008)
<i>L. peru</i>	1970-2019	CPUE (2011-2019)*	Yes	FIP monitoring in Guaymas
<i>C. princeps</i>	1980-2019		Yes	Common names reported by Manriquez-Ledezma 2008
<i>P. nebulifer</i>	1997-2019		No	-

<i>S. pulcher</i>	2000-2019	No	-
<i>S. miniatus</i>	2000-2019	Yes	FIP monitoring in El Rosario
<i>S. constellatus</i>	2000-2019	Yes	FIP monitoring in El Rosario
<i>P. auroguttatus</i>	2000-2019	No	-
<i>H. acanthistius</i>	1973-2019	No	-
<i>L. campechanus</i>	1977-2019	Yes	FIP monitoring in Nuevo Campechito

* CPUE was estimated from the fishing monitoring developed by the communities at the beginning of the San Cosme-Punta Coyote Corridor FIP. CPUE was estimated as catch (kg)/fishing trip.

Even if it has been proposed to differentiate stock between the west coast of the Baja California peninsula and the Gulf of California, none of the evidence is conclusive. For these analyses, all species were evaluated as a single stock from the Northwestern Mexican Pacific, including landings from the states of Baja California (BC), Baja California Sur (BCS), Sonora (SON), and Sinaloa (SIN). The red snapper in the Gulf of Mexico, a single stock for the Mexican part, was also considered for the analysis.

2.2. Landing reconstructions

Because official landings in Mexico are not reported at the species level, landing reconstructions must be made for some stocks. Reconstructions were done based on the methods described in Saldaña-Ruiz et al. [12]. For the California sheephead, the barred sand bass, the goldspotted sand bass, and the rooster hind, all landings reported in official statistics were considered to belong to the species evaluated because no other species are reported under any of their names (Table S1).

The Pacific red snapper is included in the “huachinangos and pargos” group along other nine species (Table S1). Most of these species are commonly reported under the “pargo” category, which is differentiated in official statistics from “huachinango”, which includes *L. argentrivetris*, *L. jordani* and *L. peru*. Based on the monitoring developed by the Guaymas FIP, *L. peru* contributes 92.5 % of the total catch of the “huachinango” group, so this percentage was applied to the reported official landings to estimate the reconstructed landings for this species. As with the Pacific red snapper, the yellowtail is reported along with another

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4 eight species (Table S1) under the common name “jurel” [42]. The yellowtail has a
5 distribution limited to the west coast of the Baja California peninsula and the Gulf of
6 California. Other species are distributed in southern areas [43,44], so “jurel” landings from
7 BC, BCS, SON and SIN were assumed to belong to *S. lalandi*.
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10 The ocean whitefish is reported with two other species in the group “pierna y conejo.” On
11 BC and BCS, *C. princeps* is named “pierna” or “blanco/blanquillo,” while *C. affinis* is
12 reported as “conejo” [45]. From 1980 to 1999, all “pierna” landings were assumed to belong
13 to ocean whitefish, while from 2000 to 2019, landings of “pierna” and “blanquillo” were
14 considered to belong to this species. The starry and vermillion rockfish are reported under
15 the category “rocote”. To estimate the specific landings for each species, the information
16 produced in the monitoring implemented by El Rosario’s FIP was used. The percentage for
17 *S. miniatus* (97.31%) and *S. constellatus* (1.06%) in total “rocote” landings was applied to
18 the entire landing for the group.
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22 The northern red snapper is aggregated with other two species (Table S1) under the common
23 name of “huachinango”. Based on the monitoring made by the Nuevo Campechito’s FIP, the
24 northern red snapper contributes 81.3% of total “huachinangos” landings, so this factor was
25 applied to all landings reported in official statistics from 1977 to 2019.
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28 **2.3. Stock status assessment**

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30 Due to the lack of species-specific landings and abundance indexes (CPUE) for all the
31 evaluated species, none of the stock could be assessed with traditional stock assessments, so
32 we used a data-poor approach to estimate stock status. Despite the limitations previously
33 described for landing data in Mexico, this information is the most reliable fishing data in the
34 country and has been collected for more than 50 years.
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37 The model chosen was the C_{MSY} [30] which fits a Schaefer model and applies Monte Carlo
38 simulations to produce estimations for reference points like stock size (B/B_{MSY}) and
39 exploitation rate (F/F_{MSY}) based on a catch time series and measurements of the resilience of
40 the stock (Martell & Froese, 2013). It also produces proxies for MSY, the biomass level
41 producing the MSY (B_{MSY}), and the fishing pressure related to the MSY level (F_{MSY}). When
42 an abundance or biomass index (i.e., CPUE) is available (even if temporal gaps exist), C_{MSY}
43 could apply a Bayesian state-space Schaefer surplus production model to refine the
44 estimations related to stock status [30]. The detailed framework of the model is described in
45 Froese et al. (2017). Briefly, the essential biomass dynamics used for the estimation are
46 estimated using the equation:
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$$50 \quad \mathbf{B}_{t+1} = \mathbf{B}_t + r(1 - \mathbf{B}_t/k) \mathbf{B}_t - C_t$$

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53 B_{t+1} is the biomass in the following year $t+1$, B_t is the current biomass, r is the rate of
54 population increase, k is the population's carrying capacity, and C_t is the catch in year t . This
55 model requires information on the catch, priors for r , and biomass depletion (B/k) at the time
56 series' beginning and end. Catch time series were described in the previous section. For the
57 prior ranges for r , values from resilience levels reported in FishBase [46] were used to match
58 categories reported by Froese et al. (2018). Specific values were available for the yellowtail
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and the Pacific red snapper (Table 3). For the priors of biomass at the beginning and the end of the time series, initial value ranges proposed by Froese et al. [30] were used based on existing information for the species (Table 4). All the analyzed stocks were assumed to be minimally exploited at the beginning of the time series (*low depletion*). For the end of the time series, the status reported in the National Fishing Chart was used, even if this status was not reported at species-level (see previous section). The current stock status in that document could be described as “overfished” or “harvested at the maximum limit,” which was considered analogous to *medium depletion*.

Table 3. Prior ranges for population growth rate (r) were used to assess ten stocks in Mexico based on the classification of resilience according to Froese et al. (2017). Ranges in bold are species-specific.

Resilience (r)	Prior range	Stocks
High	0.6 – 1.5	
Medium	0.2 – 0.8	<i>S. lalandi</i> (0.16-0.49), <i>L. peru</i> , <i>P. nebulifer</i> , <i>L. campechanus</i> (0.28-0.81)
Low	0.05-0.5	<i>P. auroguttatus</i> , <i>C. princeps</i> , <i>S. constellatus</i> , <i>S. miniatus</i>
Very low	0.015-0.1	<i>S. pulcher</i> , <i>H. acanthistius</i>

Table 4. Prior relative biomass (B/k) ranges used for assessing ten stocks in Mexico based on the depletion status suggested by Froese et al. (2017).

Depletion	Prior range	Stocks at beginning	Stocks at the end
Very low	0.6-1		

Low	0.4-0.8	<i>S. lalandi</i> , <i>L. peru</i> , <i>C. princeps</i> , <i>P. S. constellatus</i> <i>nebulifer</i> , <i>S. pulcher</i> , <i>S. miniatus</i> , <i>S. constellatus</i> , <i>P. auroguttatus</i> , <i>H.</i> <i>acanthistius</i> , <i>L. campechanus</i>
Medium	0.2-0.6	<i>S. lalandi</i> , <i>L. peru</i> , <i>C. princeps</i> , <i>P. nebulifer</i> , <i>S. pulcher</i> , <i>S.</i> <i>miniatus</i> , <i>P. auroguttatus</i> , <i>H.</i> <i>acanthistius</i> , <i>L. campechanus</i>
Strong	0.01- 0.4	
Very strong	0.01- 0.2	

The fishing cooperatives involved in the FIPs have developed monitoring programs for the species evaluated. This monitoring is a recent effort (<five years), so the information on abundance indexes could not be used to improve the estimations of the C_{MSY} . However, for the yellowtail and the Pacific red snapper, CPUE produced by other FIP in the region (Table 1) were used to include them in the assessment. In addition, Kobe plots were constructed to have a better visualization of stock status. All analyses were performed in R software.

3. Results

3.1. Reconstructed landing time series

Between 1974-1988, yellowtail landings fluctuated between 773 t and 1,989 t (Fig. 2). After that, catches declined to less than 500 t in 1992, followed by a steady increase until reaching a maximum peak between 2010-2014 (~2,800 t). First, most of landings were reported in BC, but in the eighties, landings from BCS surpassed them, contributing between 75–85% of total landings for the stock in the last ten years (Fig. S1). Guaymas, where the FIP is based, was the fishing office that contributed more to the state landings of SIN between 2000-2019, with an average of 69% (Fig. S2). The FIP contributed less than 1% of total landings for the stock in 2019 (Table 5).

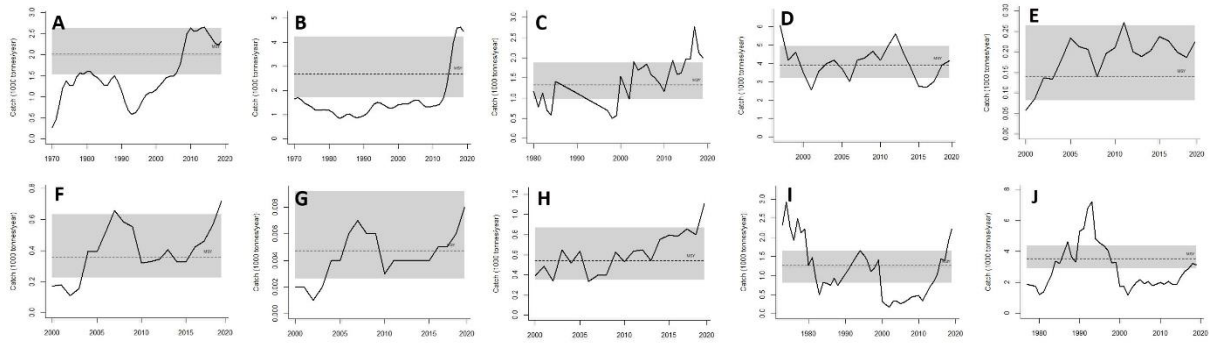


Figure 2. Catch trends for ten stocks in Mexico. The solid line indicates catches, the discontinued line indicates the catch in the reference point (C_{MSY}), and the gray area is the CI 95%. A = *S. lalandi*. B = *L. peru*. C = *C. princeps*. D = *P. nebulifer*. E = *S. pulcher*. F = *S. miniatus*. G = *S. constellatus*. H = *P. auroguttatus*. I = *H. acanthistius*. J = *L. campechanus*.

For the Pacific red snapper, reported landings were below 2,000 t until 1980 (Fig. 2), with a higher contribution of SON landings (Fig S3). From 1980 to 2014, landings ranged between 425 and volumes slightly above 1,000 t (Fig. 2), with the majority coming from BCS (Fig. S3). An increase above 2,000 t occurred in 2015, and since then, landings have triplet to reach a maximum peak in 2017 (4,067 t). Most landings were reported in BCS (~55% in 2019, Fig. S1). Between 2000-2019, most landings in SON came from Huatabampo (almost 80% of total state landings), with Guaymas contributing less than 10% (Fig. S4). In 2019, the FIP contributed 0.16% of total landings from the stock (Table 5).

Table 5. Total landings and landings reported by the communities involved in the Fishery Improvement Projects for ten stocks in Mexico in the last year of the time series (2019). Percentages of landings relative to the total stock landings are shown in parentheses. NI = Natividad Island, ROS = El Rosario, GUY = Guaymas, NCAM = Nuevo Campechito.

Species	Total catch (t)	NI (t)	ROS (t)	GUY (t)	NCAM (t)
<i>S. lalandi</i>	2,084.4	-	-	14.6 (0.70%)	-
<i>L. peru</i>	2,920.7	-	-	4.66 (0.16%)	-
<i>C. princeps</i>	1,998.7	1.43 (0.07%)	32.98 (1.65%)	0.03 (<0.01%)	-

<i>P. nebulifer</i>	4,139.7	-	8.70 (0.21%)	-	-
<i>S. pulcher</i>	223.8	-	1.04 (0.46%)	-	-
<i>S. miniatus</i>	717.1	-	173.38 (24.18%)	-	-
<i>S. constellatus</i>	7.8	-	1.88 (24.01%)	-	-
<i>P. auroguttatus</i>	1,107.7	-	-	23.73 (2.14%)	-
<i>H. acanthistius</i>	2,216.0	-	-	4.6 (0.21%)	-
<i>L. campechanus</i>	3,109.71	-	-	-	61.92 (1.99%)

From 1980 to 1985, landings of the ocean whitefish were <1,410 t (Fig. 2). Between 1986 and 1996, landings for this species were absent in official statistics. Since 1997, landings increased constantly to a maximum peak in 2017 (2,736 t) with landings averaging 2,100 t in the last five years (Fig. 2). During all the time series, most landings (>90%) were reported in BCS (Fig. S5). From 2000-2019, <2% of total landings of BCS came from the Fishing office where the FIP of Natividad Island reports their landings (Bahía Tortugas, Fig. S6). In the last six years, landings from Guaymas had the highest contribution to SON total landings (Fig. S7). Among the three FIPs catching this species, the one from El Rosario contributed 1.65% of total landings from the stock in 2019, followed by Natividad Island (0.07 %) and Guaymas (<0.01%, Table 5).

For the barred sand bass, reported landings were low until 1982 (<440 t) and then disappeared from official statistics until 1997. In that year, the highest peak of landings was reported (6,073 t), followed by fluctuating landings (2,544-5,600 t) until 2019 (Fig. 2). Most landings were reported in BCS (>90% of total landings), with minimal contribution from SIN and SON (Fig. S8). In the latest five years, the office of El Rosario, where the FIP is located, recorded the highest landings within BC (Fig. S9), contributing 0.21% of total landings for the stock in 2019.

Landings for the California sheephead appeared in yearbooks from 1972 to 1974 and disappeared from official statistics until 2000. Since that year, landings have ranged between 58 t and a maximum peak of 270 t in 2011 (Fig. 2). Most landings came from the west coast of BC and BCS (Fig. S10). In BC, the highest landings came from the Fishing office of El Rosario from 2000-2019, contributing up to 70% of total state landings (Fig. S11). The

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4 contribution to total landings from El Rosario's FIP for this species was meager in 2019
5 (0.46%, Table 5).
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8 For the vermilion rockfish and starry rockfish, landings trends follow the same pattern
9 because both species are reported under the same common name, "rocot" (Fig. 2). Landings
10 for "rocot" were sparse until 1999. Since 2000, landings for the vermilion rockfish have
11 increased until reaching a first peak in 2007 (658 t) followed by a decrease until 2015. In the
12 last three years, landings increased to a maximum peak (717 t). For the starry rockfish, the
13 highest estimated landings were found in the last year of the time series (8 t). Between 2000-
14 2019, most landings (58-90%) were reported in BC (Fig. S12), with some years representing
15 >90% of total landings for the stock. From 2000-2009, El Rosario was the office with the
16 highest landings for both species in BC (45%) and since then it has contributed between 30-
17 46% of total state landings (Fig. S13). In 2019, the FIP from El Rosario contributed 24.18%
18 and 24.01% of total landings for the stock of *S. miniatus* and *S. constellatus*, respectively
19 (Table 5).
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23 Landings for the golspotted sand bass were available since 2000 (Table 2). Until 2012,
24 catches fluctuated below 700 t (Fig. 2). Since then, landings increased until reaching a
25 maximum peak in 2019 (1,297 t). Most landings were reported in SON, representing between
26 30-50% of total landings with the contribution of BC and BCS increasing in the later years
27 (Fig. S14). Later, landings from BC and BCS increased to levels like those from SON
28 (approximately 30%), but in the last year, landings from SON contributed 50% of total
29 landings from the stock. From 2000-2002, most landings in SON came from the fishing office
30 of Guaymas (Fig. S15), which contributed between 4-12% of total state landings in the last
31 five years (Fig. S15). In 2019, the landings from the Guaymas FIP represented 2.14% of total
32 stock landings (Table 5).
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36 The rooster hind had the highest peak of landing at the beginning of the time series (2,981 t),
37 followed by a decrease until 1988 (741 t, Fig. 2). Between 1989 and 1993, reports of landings
38 in the yearbooks were absent. From 1994 to 1999, reported landings ranged between 1,100-
39 1,650 t, followed by a decrease until 2011, with landings below 500 t. Since 2012, a steady
40 increase in landings started until the end of the time series, reaching 2,200 t (Fig. 2). Landings
41 were reported mainly in SON (Fig. S16). Between 2000-2019, landings reported in Guaymas
42 ranged between 7 and 113 t, representing up to 75% of the state total in the first three years
43 and currently contributing 11% (Fig. S17). The contribution of the Guaymas FIP to total
44 landings in 2019 was 0.21% (Table 5).
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48 Landings from the northern red snapper increased since the beginning of the time series until
49 a maximum peak in the early nineties (7,200 t; Fig. 2). Until 1999, landings were reported in
50 a higher proportion in Yucatán, followed by Veracruz and Tamaulipas (Fig. S18). Following
51 that year, landings decreased to a minimum catch in 2003 (Fig. 2). Between 2004 and 2015,
52 landings were constant at around 2,000 t, with most landings reported in Tabasco (Fig. S18).
53 Since 2015 a slight increase was detected, surpassing 3,000 t in 2019, with most landings
54 reported in Campeche (Fig. 2, S18). The Atasta office, where the FIP reports their landings,
55 contributed between 0.69% and 14.76% of total state landings between 2000 and 2019 (Fig.
56 S19). In 2019, the landings from the FIP of Nuevo Campechito represented 1.99% of total
57 landings from the stock (Table 5).
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3.2. Stock status assessment

From the ten stocks evaluated, the relative biomass in the last year of the time series is above the level producing MSY ($B/B_{MSY} < 1$) in four stocks (Fig 3, Table 6). In another three, relative biomass is below MSY but around the reference point used by the MSC (0.9 MSY). In the remaining three, the stock biomass is below the reference point but above the point where the recruitment would be impaired ($0.5 B_{MSY}$, Fig. 3). The biomass of the rooster hind stock has been below MSY the whole time series.

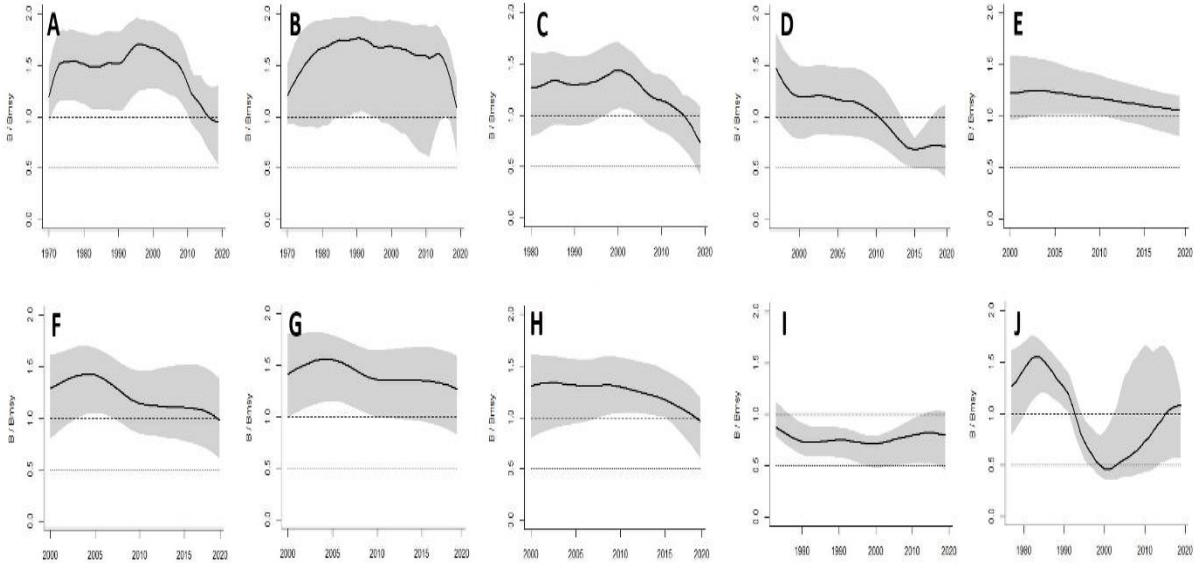


Figure 3. Relative biomass trends (B/B_{MSY}) for ten stocks in Mexico. The discontinued line indicates the biomass in the reference point (B_{MSY}), the pointed line indicates the level where the recruitment could be impaired ($0.5 B_{MSY}$), and the gray area is the uncertainty. A = *S. lalandi*. B = *L. peru*. C = *C. princeps*. D = *P. nebulifer*. E = *S. pulcher*. F = *S. miniatus*. G = *S. constellatus*. H = *P. auroguttatus*. I = *H. acanthistius*. J = *L. campechanus*.

Table 6. Estimates of biological parameters and reference points for ten stocks in Mexico. Reference points presented are biomass in the reference point (B_{MSY}), relative biomass in the last year (B/B_{MSY}), fishing mortality in the reference point (F_{MSY}), and fishing pressure in the previous year (F/F_{MSY}). Values in parentheses belong to the estimates without incorporating the CPUE index.

Species	r	K (10^3 t)	B_{MSY} (10^3 t)	B/B_{MSY}	F_{MSY}	F/F_{MSY}
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<i>S. lalandi</i>	0.32 (0.33)	24.94 (23.46)	12.47 (11.73)	0.95 (0.70)	0.16 (0.16)	1.21 (1.77)
<i>L. peru</i>	0.40 (0.46)	26.75 (25.26)	13.37 (12.63)	1.08 (0.94)	0.20 (0.23)	1.53 (1.42)
<i>C. princeps</i>	0.22	24.65	12.32	0.73	0.11	1.95
<i>P. nebulifer</i>	0.60	26.52	13.26	0.71	0.30	1.44
<i>S. pulcher</i>	0.05	11.35	5.67	1.06	0.03	1.41
<i>S. miniatus</i>	0.23	6.85	3.43	0.98	0.12	1.84
<i>S. constellatus</i>	0.22	0.08	0.04	1.27	0.11	1.24
<i>P. auroguttatus</i>	0.22	10.43	5.21	0.96	0.11	1.95
<i>H. acanthistius</i>	0.03	127.10	63.55	0.79	0.01	2.22
<i>L. campechanus</i>	0.38	37.88	18.94	1.07	0.19	0.80

Regarding the fishing pressure, all stocks are currently at levels above the reference point ($F/F_{MSY} > 1$), except for the northern red snapper ($F/F_{MSY} = 0.80$) and have been at this level for the past three years (Fig. 4, Table 6). In almost all stocks, the fishing pressure was at levels below MSY ($B/B_{MSY} < 1$) for most of the time series (Fig. 4). However, this reference point was surpassed in the last ten years because of the increases in catches described in the previous section. Estimations from r and k were under the limits set as a prior (Fig. S20).

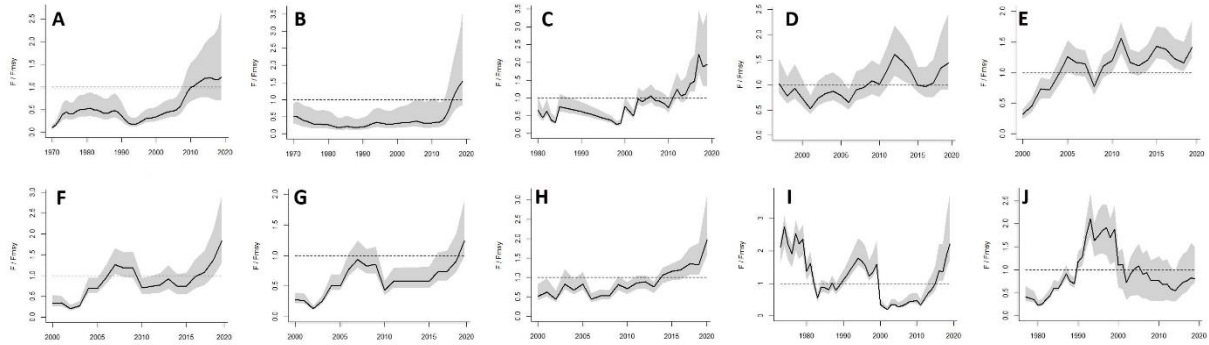


Figure 4. Relative fishing pressure trends (F/F_{MSY}) for ten stocks in Mexico. The discontinued line indicates the fishing mortality in the reference point (B_{MSY}), and the gray area is the uncertainty. A = *S. lalandi*. B = *L. peru*. C = *C. princeps*. D = *P. nebulifer*. E = *S. pulcher*. F = *S. miniatus*. G = *S. constellatus*. H = *P. auroguttatus*. I = *H. acanthistius*. J = *L. campechanus*.

According to the Kobe plots, there is a high probability that six of the stocks are currently overfished with overfishing occurring (Fig. 5). For the other three stocks, overfishing is happening, and only the stock of the northern red snapper is in optimal conditions. As well as with the biomass and fishing pressure trends, the stock status of these species changed drastically in the last ten years of the time series due to increases in landings (Fig. 2).

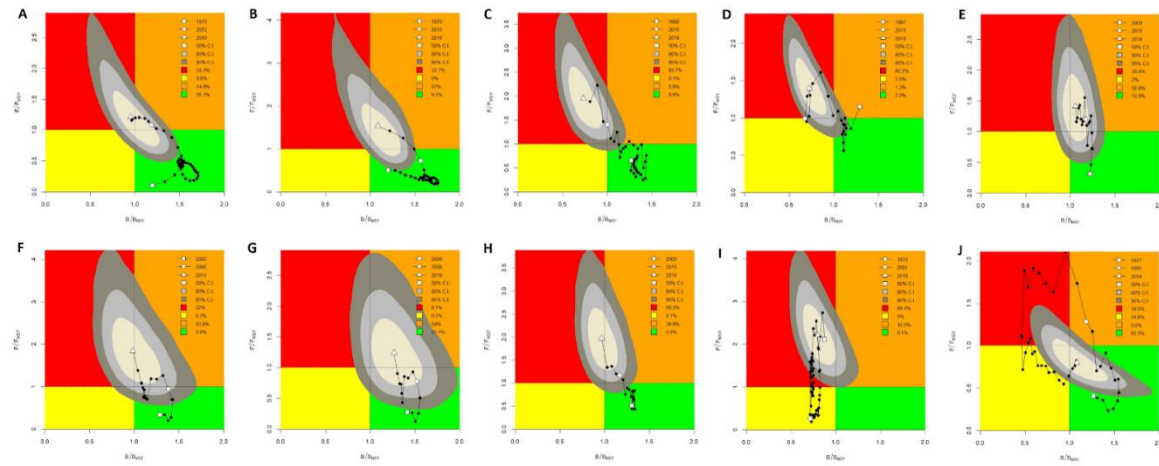


Figure 5. Kobe plots with the stock status trend for ten stocks in Mexico. The gray area around the triangle indicates uncertainty related to the final year (yellow for 50%, grey for 80%, and dark grey for 95% confidence levels). Legend indicates the probability of the stock status being of each of the plot quadrants. A = *S. lalandi*. B = *L. peru*. C = *C. princeps*. D = *P. nebulifer*. E = *S. pulcher*. F = *S. miniatus*. G = *S. constellatus*. H = *P. auroguttatus*. I = *H. acanthistius*. J = *L. campechanus*.

4. Discussion

4.1 Stock status of case studies and reconstructed landing trends

Our estimations showed that most stocks are fished at unsustainable levels, with overfishing occurring in almost all the stocks and some under overfished conditions. Except for two stocks (rooster hind and northern red snapper), this status is related to increases in landings (triplicated for some species) for the past ten years. Despite the limitations of catch-only models for stock assessments, trends in stock biomass and fishing pressure are evident: management actions are urgently needed for all species analyzed. Even if results must be taken with care due to the intrinsic limitations of this kind of model [29,47], they can be used as a baseline to assess the stock status of these fisheries.

Catch-based models, such as C_{MSY} , have proven to help evaluate the stock status of marine resources [32,48]. When these models are implemented with informative priors, their performance in estimating the stock status improves, making them a helpful tool for evaluating fisheries in areas where robust stock assessments are unavailable [48]. Some cautions need to be considered when using this model type. Pons et al. [49] found that catch-based model performance highly depends on life-history parameters (i.e., r , K), depletion levels, and fishing intensity. Nevertheless, catch-only models can give good results for referent points like MSY when long catch-time series are used [29]. For this reason, through the reconstruction process of the catches in this study, we tried to have the most extended landing time series for all the stocks analyzed.

Increases in landings for the yellowtail in BCS were reported previously [50], which could be related to increases in fishing effort between 2005-2011 [51]. The historical fishing importance of the Pacific red snapper in the Gulf of California has increased over the years, gaining significant relevance in the 1990s and 2000s (Sala et al., 2004). Our time series includes landings since 1977 but older records indicate high economic relevance and local consumption of snappers in the Gulf of California since the 1940s [52], which were not found in Official Fishing Statistics nor included in the analyses. Sand bass fishing has been highlighted for the Gulf of California since the 1940s and the mid-1990s [52,53], but official statistics only reported landings since 2000, when an increase in catches was previously reported [54].

The ocean whitefish fishery in the Northwest Mexican Pacific can be traced back to the 1940s [52], and until the 1980s, it was considered a low-importance species [55], which may explain their absence from official landings between 1986-1996. As our catch reconstruction shows, its importance in landings increased in the late nineties [55]. In the Gulf of California, an intense fishing effort for the rooster hind was reported in coastal SON and BC during the seventies [53]. In the eighties, rooster hind catches decreased due to the rising demand for shark meat in Mexico until it resurfaced in the nineties [53,54], followed by a decrease until 2010, as our data shows. For the California sheephead, official landings were found in the early seventies and then disappeared from official statistics until 2000, around the time when

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4 its fishing importance in Baja California, Mexico, and Southern California, US, was reported
5 [56–58]. Like our catch reconstruction, the fishery of the barred sand bass on the Pacific coast
6 of BC and BCS became important in 2001 [59], which could explain its presence in official
7 fishing statistics until late 1990s [60]. The vermillion and starry rockfishes have been
8 commercially fished in California since the nineteenth century [61,62], but in the Baja
9 California peninsula has been reported as important since the 2000s [57] due to the lack of
10 studies on the history of the fishery in the region.
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14 The catch of the northern red snapper can be traced to the late 1890s in Campeche Bank,
15 being this region one of the most important between 1950-1970 [63]. Estimated landings
16 decrease trends of this species in the southern Gulf of Mexico during the nineties have also
17 been reported by Perez-Jimenez et al. [13]. As our results show, FIPs and the promoted
18 citizen science can be valuable tool to improve catch reconstructions that provide insights
19 into the fishery trends of the species and allows us to step further in future population
20 assessments and the development of management strategies.
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24 Biomass from most of the evaluated stocks was above the reference point in most time series,
25 except for the rooster hind and the northern red snapper. In the last years, the stock biomass
26 started to decline to levels around and below the reference point. These declines are related
27 to an increase in landings in the past ten years that has also been reported for other fishing
28 resources in Mexico [11,18,64]. Due to deficiencies in the fishing report system in the
29 country, it is impossible to identify if this increase is related to changes in fishing effort. Even
30 in stocks that were assessed with shorter time series (e.g., barred sand bass) these declines in
31 biomass have occurred in the last five years. For the rooster hind, the largest catches were
32 identified at the beginning of the time series producing stock sizes below MSY and
33 continuing at a constant level for the rest of the time series. Besides the decrease in landings
34 for this species, another reason the stock biomass did not decrease at lower levels could be
35 that this species has a larger stock size (the largest K estimated among all stocks). The
36 northern red snapper's stock biomass decreased below MSY in the early nineties. This stock
37 was reported as overexploited in Mexico [10], with a population reduction of around 58%
38 [65]. In our estimations, the stock exceeded the reference point in the last three years after
39 overexploiting. Like the stock biomass, the increase in landings for the past ten years has
40 produced unsustainable fishing pressure for all the stocks except for the northern red snapper.
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46 The lack of data in Mexican fisheries is a common feature preventing the development of
47 formal stock assessments needed to implement management strategies [23,66]. An
48 alternative data source is the logbooks filled out by fishing cooperatives or fishing permit
49 owners, even if this source data still has some limitations (e.g., [67,68]. Fulton et al. [69]
50 recommended the institutionalization and adoption of citizen science by fisheries
51 management agencies and researchers to help create national data collection networks.
52 Therefore, to improve the stock status of Mexican fisheries and in line with recommendations
53 from other studies [12,70], we propose some considerations: 1) although landings
54 reconstruction methods have proven to diminish uncertainty on species-specific catch records
55 in data-poor areas and this could be improved by citizen science, these methods do not
56 exclude the need to have official catch records at the species level; 2) efforts should be made
57 to estimate updated life-history parameters for exploited stocks to increase dynamic
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4 population knowledge (i.e., size at first maturity, growth parameters, estimate natural, and
5 fishing mortality), and 3) detailed information of fishing effort (i.e., fishing gears including
6 type and sizes, soaking times) associated with catches should be collected so index of
7 abundances (e.g., CPUE) could be estimated. Addressing these topics will allow the
8 implementation of more robust stock assessment. Control rules to regulate fishing efforts
9 and landings need to be established and published in specific Fishing Management Plans to
10 secure that stock biomass and fishing pressure fluctuate around the reference point.
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13 **4.2 Management implications and recommendations**

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16 In the latest version of the National Fishing Chart that included finfish fisheries, the central
17 management recommendation was to maintain the current level of captures, as the fishery is
18 exploited at its maximum sustainability level [9]. Among the finfish species considered in
19 this work, only the management fishery plan for the barred sand bass in the Baja California
20 peninsula has been published [36]. Other finfish species need management plans, harvest
21 strategies, and stock assessments, making their fisheries management weak. Based on the
22 NFC, two general management actions are possible: 1) Fishers need to own a finfish fishing
23 permit (“permiso de escama”), which encompasses ~70 fish species, and 2) if landing
24 volumes fall from a specific threshold, more management actions need to be enforced for the
25 *S. lalandi*, *P. auroguttatus*, *P. nebulifer*, *C. princeps*, *H. acanthistius*, *L. peru* [42], *S. miniatus*
26 and *S. pulcher* [9]. However, the NFC make these recommendations for categories that
27 include multiple species, preventing that management action can be defined at the species
28 level (for example, the Carangidae complex group “jureles” and “medregales” contains nine
29 species, and the Serranidae complex group includes 18 species) [42]. The California
30 sheephead is considered a bycatch in the Serranidae complex group; therefore, the NFC only
31 established the use of a finfish permit for its harvest as a management action.
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37 Although the northern red snapper is one of the leading fishing resources in the Gulf of
38 Mexico [71], the NFC established that the Fishing Management Plan needs to be elaborated
39 and implemented [10], which until this date, has not been done. As stated above, the barred
40 sand bass is the only assessed species in this study with a Fishing Management Plan [36].
41 Based on its stock status, it suggests that biomass should be maintained below B_{MSY} and
42 established the formulation of harvest control rates, minimum catch size, and regulation of
43 fishing gears. This plan was published very recently, so it is unknown if and how these
44 actions have been implemented.
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47 All the stocks evaluated are part of Comprehensive FIPs, except the ocean whitefish targeted
48 in Natividad Island, meaning an evaluation applying the MSC standard should be made as
49 part of the reporting. Based on the estimated stock status and the lack of specific harvest
50 strategies and control rules for most stocks, all stocks (except the northern red snapper) will
51 fail to achieve the highest score for most of the indicators of Principle 1 and 3. All the stocks,
52 except the rooster hind and barred sand bass, could achieve the highest score (≥ 80) for the
53 indicators related to stock status because stock sizes are around the reference points. Only
54 the barred sand bass has a Fishing Management Plan with harvest strategies and control rules
55 (even if those are not very clear), reaching a medium score for those two indicators. The rest
56 of the stocks lack these management strategies, achieving the lowest score possible. This lack
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4 of harvest strategies is critical because fishing pressure in all the stocks is above the reference
5 point, so control rules need to be established to ensure that stock sizes stay above MSY. Some
6 finfish FIPs have set internal measures such as minimum catch size or fishing effort in
7 Mexico but at a fishing community level. This effort must be implemented at the species
8 distribution level to influence population status.
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11 The catch for all the stocks in this work is monitored periodically only by the logbooks that
12 fishers submit to the Fishing Offices. However, this information is not always reported at a
13 specific level and has no information on fishing effort. Under the current context, the
14 monitoring indicator under the FIP will achieve a medium score. As part of the FIPs
15 implementation, communities have developed monitoring programs for fishing and
16 biological information. This monitoring will produce information regarding relative
17 abundances by fishing gear (CPUE) and specific landings that, in the short term, will aid in
18 improving the current stock assessment to get a more accurate estimation of stock status, like
19 with the yellowtail and the Pacific red snapper. However, this monitoring should be extended
20 to other areas where the stocks are caught, especially where the highest landings are reported.
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25 FIPs can help facilitate the organization and management of fisheries at local level. Involving
26 more communities and fishing cooperatives in the FIPs is critical to gather the amount of
27 data needed to strengthen stock status estimations needed to produce proper management
28 actions that secure the sustainable use of fishing resources. This joint work will help to have
29 a systematized information exchange between them to meet the standards and key indicators
30 on a larger scale and reach the objectives of the FIPs. However, monitoring capacity is
31 limited, and it is equally crucial that stakeholders take responsibility for obtaining this data
32 with more extensive coverage. Therefore, data collection should not be considered
33 complementary or only performed to meet MSC standards. Still, FIPs can support and
34 reinforce this critical task to seek the sustainability of the fishery [72]. Initiatives raised from
35 the bottom of the fishing system may be the key to internally establishing effective
36 management strategies, even if Fishing Authorities do not produce them.
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41 It is essential to mention that the catch from FIPs contributes less than 5% to the total yield
42 of the stock, except for the rockfish species, which has implications for fishery management,
43 as it largely depends on the population status and its assessment due to data collection and
44 quality. It is necessary to have information on a representative percentage of the total
45 population stock for a more robust evaluation. To comply the principle 1 of the MSC,
46 sustainability of the stock, the cooperatives that participate in FIPs play an essential role in
47 generating this information to fill existing gaps and contribute to the improvement of
48 practical management actions. There are examples of how fishers have provided information
49 to detect shifting fishing seasons and catch compositions, allowing them to inform
50 management decisions quickly and efficiently (Fulton et al., 2019). Citizen science-based
51 information can be used to develop strategic management plans and comprehensive models
52 despite the scarcity of data [73].
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57 To improve fishery practices toward sustainability, implementation of FIPs has been
58 increasing worldwide over the last few years [74]. Even Samy-Kamal [74] suggests that only
59 half of the implemented FIPs have improved. The FIPs we analyzed have implemented
60 systematic monitoring of their fishing operations that, in the future, will allow having catch
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4 and fishing effort time series with species compositions. Moreover, combined with the
5 necessary biological knowledge of the assessed species, it will permit the application of more
6 robust stock assessment models, either for single or multiple species [22,75]. For the case of
7 Mexico, this tool sounds very promising for a bottom-up reconciliation route for a gain of
8 regional multi-specific artisanal fisheries management. Nevertheless, deepening other
9 aspects considering the fisher's operations to understand the dynamics of the fishing effort
10 [12] and social, economic, and human constraints [76,77] might also need to strengthen
11 regional sustainable fisheries practices in Mexico.
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15 16 17 **Acknowledgments** 18

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22 snapper FIP, and Gulf of California grouper, snapper, triggerfish & yellowtail FIP. Also, we
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